

UNIVERSITY OF TARTU
INSTITUTE OF GEOLOGY

Kadri Haamer

**Hydrochemistry and sources of fluoride in Silurian-Ordovician
aquifer system**

MSc thesis

supervisor: PhD E. Karro

TARTU 2006

CONTENTS

Annotation	3
Introduction	4
Materials and methods	5
General geochemistry of fluorine	7
Health effects of fluoride	11
Geological setting	14
The Silurian-Ordovician aquifer system	17
Fluoride contents in Silurian-Ordovician aquifer system	18
Vertical distribution of fluoride	20
Hydrochemistry of fluoride	25
Possible source of fluoride-rich groundwater	32
Conclusions	37
Acknowledgements	39
References	40
Summary in Estonian	45
Appendix	47

ANNOTATION

Haamer, K. 2006. Hydrochemistry and sources of fluoride in Silurian-Ordovician aquifer system. University of Tartu, Institute of Geology. MSc thesis. 58 p.

The purpose of the current MSc thesis is to give an overview of the fluoride content in the Silurian-Ordovician aquifer system. The study aims to find the connections between environmental factors, the concentration of fluoride and other chemical elements in groundwater, as well as to explain the importance of the chemical type of groundwater as the controller of the fluoride content.

In order to study the fluoride content of the aquifer material twenty-five limestone, dolomite and K-bentonite samples were analysed. The high fluoride concentrations are well preserved in K-bentonite beds and therefore contribute to the formation of F-rich groundwater in the Silurian-Ordovician aquifer system.

Key words: Estonia, fluoride, groundwater, K-bentonite.

INTRODUCTION

Fluorides in the environment are of public and scientific interest because of their effects on health. Small doses of fluoride have beneficial effects on the teeth by hardening the enamel and reducing the incidence of caries, but excessive intake of fluoride results in dental and skeletal fluorosis. Unlike many other elements, a large portion of fluoride (F) is ingested from drinking water.

According to the WHO Guidelines for Drinking Water Quality (WHO, 2004) and EU Directive 98/83/EC (1998), as well as the Estonian Requirements for Drinking Water Quality (Joogivee, 2001) the limit value for fluoride is 1.5 mg/L.

Only few studies on the occurrence and distribution of fluoride in drinking water have been carried out in Estonia (Kuik, 1964; Saava et al., 1973; Karro & Rosentau, 2005). Unfortunately, these former studies have focused on particular region or city. On the other hand, health problems arising from different fluoride intake have been recorded (Kiik, 1973; Saava, 1998; Russak et al., 2002). There was thus a need for a study examining the spatial distribution and identifying probable natural sources of fluoride in drinking water. The results of the current study should help to work out the strategies for safe drinking water supply in case of excessive fluoride content.

The purpose of the present MSc thesis is to give an overview of the fluoride content in the Silurian-Ordovician aquifer system and to determine the areas, where the concentration of this element in groundwater is higher than the limit value set by the ministry of social affairs' s regulation. The study aims to find the connections between environmental factors, the concentration of fluoride and other chemical elements in groundwater, as well as to explain the importance of the chemical type of groundwater as the controller of the fluoride content. The dependence of abstracted groundwater chemistry on the construction of drilled wells has been studied and the possible geological sources of the high fluoride content in groundwater are also discussed.

MATERIALS AND METHODS

Current study is based on the groundwater monitoring data collected during years 1964 to 2005 and on the existing information obtained from Geological Survey of Estonia, local environmental departments and Institute of Geology, University of Tartu. The gathered groundwater chemistry data include pH, anions (HCO_3^- , SO_4^{2-} , Cl^-), cations (Na^+ , Ca^{2+} , K^+ , Mg^{2+}), F^- and B. The location, cadastre number and the depth of the wells, year of water sampling and the names of sampled aquifer systems were included into joint groundwater database, which contains all together 430 analyses (Appendix 1).

For data processing, interpretation of the results and compilation of cross-sections, MapInfo Professional 6.5, CorelDRAW 9 and AquaChem 3.7 were used.

Twenty-five limestone, dolomite and K-bentonite samples were analysed in order to study the fluoride content of the aquifer material. Rock samples for leaching experiments were ground using agate mortar.

The samples were studied using a sequential leaching method: previously boiled 18-megohm water (Millipore, Rios, Milli-Q) and 0.5 M Tiron (3,5 Pyrocacholdisulfonic acid disodium salt) solution. At first, 500 mg of rock samples were leached in 20 mL of water for 16 hr at 24 °C using an orbital swinging shaker (GFL 3011). The leachings were performed under argon atmosphere. Secondly, the 0.5 M Tiron that is regarded to dissolve F associated with oxidic Fe and Al sites of the K-bentonites, was applied for the K-bentonite samples. 15 mL of the Tiron leachate was used for 1 hr at 80 °C (Biermans & Baert, 1977).

For total fluorine contents, 500 mg of ground samples were mixed and sintered with 3500 mg of Na_2O_2 in covered Pt-dishes (Seelye & Rafter, 1950) for 45 minutes at 460 °C. After cooling, the sintrates were let to react with 50 mL of double distilled water and dissolved stepwise using four times 3.5 mL of

6 M HCl, and made up to 100 mL in 18-megohm water. Polypropylene test flasks were used in order to avoid complexation of fluorine with boron from borosilicate glass.

The F contents were measured using a fluoride selective electrode under magnetic stirring (Orion Research, Model 920+). Total ionic strength adjustment buffer (TISAB) solution was used to decomplex the F and buffer the system. The determinations of fluoride were made under even temperature and illumination (Nicholson, 1983).

GENERAL GEOCHEMISTRY OF FLUORINE

Fluorine (F) is the most electronegative and reactive of elements, and reacts with practically all of the organic and inorganic substances. It never occurs in the elemental state in nature because its high reactivity. The average fluorine concentrations in the Earth's crust is estimated at 0.05 – 0.1 % or 500 – 1000 mg/kg and ranks 13th among the elements (Turekian & Wedepohl, 1961; Adriano, 1986; Wedepohl, 1995). It is enriched during differentiation in the late stages of crystallizing magmas as well as in the residual solutions and vapour. Consequently, it is concentrated in highly siliceous granitic and alkaline rocks, and in hydrothermal mineral deposits (Hopkins, 1977).

Fluorine may occur in limestones associated with tremolite-, actinolite-pyroxene- or with mica-chlorite-skarn formations, where F^- concentrations may reach the level of 0.4 – 1.2 %. Part of fluorine may be present in clay material admixed with weathered limestones. It can also be combined with phosphate minerals (Lahermo & Backman, 2000).

The most important fluorine minerals are fluorite (CaF_2), fluor-apatite $Ca_5(PO_4)_3F$ and apatite $Ca_5(PO_4)_3(F,OH,Cl)$. They are characteristic accessory minerals in many types of rocks. Since fluorite may contain nearly half of fluorine by weight, it contributes substantially to F concentrations in rock and water even though the mineral is present in small quantities (Correns, 1956). Pure fluor-apatite ($Ca_5(PO_4)_3F$) could contain up to 3.8 % fluorine.

Fluorine may occur also in the composition of clay minerals, micas and pyroxenes. Soils dominated by clay minerals have high sorption capacity and frequently serve as natural geochemical barriers preventing pollutants from reaching groundwater. Fluorides are adsorbed on clay minerals, where ionic exchange takes place – F^- ions partly replace OH^- groups (Hübner, 1969; Flühler et al., 1982). More than 90 % of natural F^- in soils is bound to clay particles (Lahermo & Backman, 2000).

The role of micas and amphiboles in the occurrence and mobility of fluorides in soils is enhanced by their relative readiness to disintegrate in weathering processes. Fluorides in micas are likely to be released soon after the onset of weathering. According to Korting (1972) in clay fractions 80 – 90 % of fluorides are contained in muscovite, illite and related minerals of the mica group.

Apart from natural sources, a considerable amount of fluoride may be contributed due to anthropogenic activities. Remarkable amounts of fluorides are transferred into agricultural soils by phosphate fertilizers. Phosphorites may contain up to 4 % of fluorine depending on its origin and contents of fluorapatite (Fleischer et al., 1974). The use of fluorocarbons, insecticides and weed killers may also contribute to local F^- concentrations in soils. In addition to agriculture part of the fluorides are derived from industrial emissions and waste waters. The steel, aluminium, glass, brick and ceramic industries use fluoride in their production process that occurs with the encounter of F-containing aerosols, dust and wastewater into the surrounding environment (Arnesen et al., 1995; Haidouti, 1995).

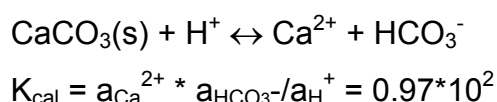
In groundwater fluorine occurs as fluoride ions (F^-) and forms soluble complexes with many kinds of inorganic and organic ligands. Fluorides are released into aqueous solution during weathering process of rocks and minerals and through anthropogenic pollution. Solubility of F from F-bearing minerals is relatively low under normal conditions, but the slow kinetics often enhances leaching and F enrichment in groundwater (Hem, 1989). Part of fluorides may occur in groundwater as a result of mineral fluorite (CaF_2) dissolution. The solubility product for fluorite (K_{fluor}) can be determined according to the following dissociation formula (Helgeson, 1969):



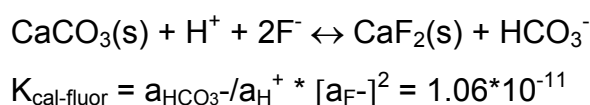
$$K_{fluor} = [a_{Ca^{2+}}] * [a_{F^-}]^2 = 10^{-9.04}, \quad \text{at } 25^\circ C$$

where $[a_{Ca^{2+}}]$ and $[a_{F^-}]$ are the activities of the concerned ions. Both Smyshlyaev and Edeleva (1962) and Handa (1975) have reported for K_{fluor} the values of $10^{-10.75}$. Due to the relatively low solubility of fluoride, the occurrence of aqueous F^- is predominantly controlled by the availability of free Ca^{2+} -ions in water (Jacks, 1973).

The dissolution constant of calcium carbonate ($CaCO_3$) is much larger than that of fluorite:



Since the solubility product for fluorite is constant, the activity of fluoride is directly proportional to HCO_3^- , if the pH is constant. When computing the thermodynamic equilibrium in the groundwater system in contact with both calcite and fluorite, a combined equation is derived (Handa, 1975):



The above values are taken from Helgeson (1969). The corresponding value by Smyshlyaev and Edeleva (1962) is $K_{cal-fluor} = 3.6 * 10^{-12}$. From the last equation can be concluded that the aqueous F^- concentrations are proportional to HCO_3^- concentrations and pH values. Consequently, highly fluorine waters are usually HCO_3^- -dominating waters what favour the dissociation of fluoride from soils. Therefore, water with high concentrations of fluorides can form in the areas where alkaline carbonate-containing waters are in contact with fluorine-bearing rocks.

Fluoride concentrations are relatively independent of the other water-soluble components, but noteworthy correlation exists between F^- and pH values. The F^- solubility in soils is lowest in the pH range of 5.0-6.5 (Adriano, 1986). At higher pH values ionic exchange occur between F^- and OH^- ions (illites, chlorites, micas and amphiboles) (Englund & Myhrstad, 1980) that leads to

fluoride increase in groundwater. In acid environment ($\text{pH} < 6$) both F^- and Al^{3+} steep into water and formation of $[\text{AlF}]^{2+}$ and $[\text{AlF}_2]^+$ complexes mainly takes place in solution (Hem, 1989; Wenzel & Blum, 1992; Haidouti, 1995; Neal, 1995). Hitchon (1995) has observed that the formation of $[\text{MgF}]^+$ complexes on the account of F-ions, intensifies while the saltiness of groundwater, Ca^{2+} content and temperature increases.

During the volcanic eruptions a lot of volcanic gas and ash fall into atmosphere and to ground from there. Plumes of ash and dust, which may cover hundreds of kilometres, may carry a variety of adsorbed chemicals. After the ash settles to the ground, these toxins may be dissolved and thus leach into the environment. The dominant chemicals that may adsorb to tephra and thus act as leachates are: chlorine, sulphur compounds, sodium, calcium, potassium, magnesium and fluorine (Weinstein & Cook, 2005). Other elements are present in smaller quantities. Various studies of volcanic emission have shown that volcanic gases and ash are rich in fluorine (Barclay et al., 1996; Cronin et al., 2003), thus clay-rich volcanic ash beds provide adsorption and ion exchange sites for F-ions. Fluorine occurs mainly in the form of HF, H_2F_2 or F_2 as well as SiF_4 and H_2SiF_6 (Fleischer et al., 1974).

HEALTH EFFECTS OF FLUORIDE

Fluorides in the drinking water are in the focus of public and scientific interest because of their effects on health. Unlike many other elements, a large portion of fluoride is ingested from drinking water. The optimal value for fluoride in drinking water is considered to be around 0.8-1.5 mg/L (WHO, 2004), whereas both deficiency and excess of F^- might be harmful from the dental point of view. A higher concentration of F^- causes dental fluorosis, and a concentration less than 0.8 mg/L results in dental caries. Fluoride makes the tooth structure stronger, so teeth are more resistant to acid attacks. Acid is formed when the bacteria in plaque break down sugars and carbohydrates from the diet. Repeated acid attacks break down the tooth, which causes cavities. Fluoride also acts to repair, or remineralize, areas in which acid attacks have already begun. The remineralization effect of fluoride is important because it reverses the early decay process as well as creating a tooth surface that is more resistant to decay (ADA, 1999). Therefore in the areas with fluorine-rich drinking water the damage of caries to teeth is significantly smaller compared with the areas with fluoride-deficient drinking water. Table 1 illustrates the different tooth decays according to the content of fluoride in drinking water.

Table 1. Relation between tooth decays and fluoride content in the drinking water (Kuik, 1964).

F^- content in drinking water (mg/L)	Tooth damages
up to 0.5	Wide spread of dental caries.
0.6-0.8	Dental caries appears often, but not so widely.
0.9-1.2	Dental caries spreads narrowly. No fluorosis.
1.3-1.5	Dental caries spreads narrowly. Rarely mild fluorosis.
1.6-2.0	Dental caries spreads narrowly. Mild fluorosis.
above 2.0	Spread of dental fluorosis.

Dental fluorosis is the damage of tooth enamel, which is caused by the long-time consumption of water with high fluoride content during the period of development of the tooth. After tooth enamel is completely formed, dental

fluorosis cannot develop even if excessive fluoride is ingested (WHO, 1984). Hence, older children and adults are not at risk for dental fluorosis. Teeth impacted by fluorosis have visible discoloration, ranging from white spots to brown and black stains. Teeth with fluorosis also have an increased porosity of the enamel. In the milder forms, the porosity is mostly limited to the sub-surface enamel, whereas in the more advanced forms, the porosity impacts the surface enamel as well, resulting in extensive pitting, chipping, fracturing, and decay of the teeth (Russak, 1996; Altosaar et al., 1998). Dental fluorosis has been classified in a number of ways. Dean (1934) developed one of the most universally accepted classifications, where Dean's index measures the degree of mottled enamel (fluorosis) in teeth. According to the damage of tooth very mild, mild, moderate and severe fluorosis can be defined (Table 2). Long-term ingestion of large amounts of fluoride can lead to potentially severe skeletal problems, such as osteosclerosis - an abnormal hardening or increased density of bone tissue. Links between high fluoride and other health problems, including birth defects (Hamilton, 1992) and cancer (Marshall, 1990) are less clearly defined.

Table 2. Degree of difficulty of dental fluorosis (Dean, 1934).

Dean's index	Tooth damages
I	Very mild fluorosis: small opaque, paper white areas scattered irregularly over the tooth but not involving as much as 25% of the tooth surface.
II	Mild fluorosis: the white opaque areas in the enamel of the teeth are more extensive but do not involve as much as 50% of the tooth.
III	Moderate fluorosis: all enamel surfaces of the teeth are affected, and the surfaces subject to attrition show wear. Brown stain is frequently a disfiguring feature.
IV	Severe fluorosis: all enamel surfaces are affected and hypoplasia is so marked that the general form of the tooth may be affected. The major diagnostic sign of this classification is discrete or confluent pitting. Brown stains are widespread and teeth often present a corroded-like appearance.

Many high-fluoride groundwater provinces have been recognized in various parts of the world. High fluoride contents, up to 40 mg/L have been reported in China, India, Africa, Japan and Mexico (Agrawal & Vaish, 1998; Lahermo et al., 1991; Yong & Hua, 1991; Alarcón-Herrera et al., 2001). Fluoride removal

practices vary widely and many high-fluoride water sources are used without treatment. As a result, large populations throughout the world suffer the effects of chronic endemic fluorosis.

The concentration of fluoride in Estonian drinking water varies widely, from 0.01 mg/L to 7.2 mg/L (Karro et al., 2006). As a result of that the extent of health problems, like dental caries and dental fluorosis must be studied in detail. The results of some earlier studies performed in Estonia are shown in Table 3. There exists a clear relation between fluoride content and the degree of occurrence of dental fluorosis and caries.

Table 3. Incidence of dental caries and fluorosis depending on the fluoride content in drinking water (Kiik, 1973; Saava, 1998).

Fluoride content in drinking water (mg/L)	Prevalence of caries (%)	Prevalence of fluorosis (%)
0.1-0.5	82.6	
0.6-1.0	70.7	6.3
1.1-1.5	59.1	11.2
1.6-2.0	48.6	38
2.1-2.4	50.9	41.9
2.5-2.8		54.5
2.9-5.0		88.9

GEOLOGICAL SETTING

The Ordovician rocks are exposed in a narrow (40-50 km) belt extending from the west to east for about 300 km in Estonia. The practically undisturbed strata are almost flat-lying with a small dip to the south. The sequence consists of highly fossiliferous shallow water open shelf carbonates – clayey limestones and marlstones, with the exception of its basal part which is represented by terrigenous sediments: sandstones, clays and graptolite argillites (Table 4). The total thickness of the Ordovician varies from 70 to 180 m, being maximal in the central and eastern Estonia (Rõõmusoks, 1983).

Two main facies belts of the Ordovician Basin are distributed in Estonia: the marginal and the central belt. Shallow water carbonate sediments with a lot of discontinuity surfaces dominated in the marginal confacies belt, while the relatively deeper water central belt comprised predominantly clayey sediments such as clayey limestones and marlstones (Nestor & Einasto, 1997). The Ordovician System of Estonia includes all three global Series: Lower-, Middle- and Upper-Ordovician and eighteen regional Stages are used as regional stratigraphic standard for Baltoscandian area (Table 4).

During the Silurian, relatively shallow water sedimentary conditions prevailed in Estonia. According to the facies model worked out by Nestor and Einasto (1997), five facies belts can be distinguished in the Silurian Basin: tidal flat/lagoon, shoal, open shelf, transition and basin depression. The first three facies belts formed a carbonate shelf or carbonate platform and the latter two - a deeper basin with fine-clastic deposits. During the Silurian shallow marine facies belts generally shifted in basinward direction as a result of basin infilling, but local conditions changed in course of sea level changes (Nestor & Einasto, 1997). The composition of sediments (mainly clay content) reflects the sea level changes of the Silurian Basin. The Silurian transgression maximum coincides with Juuru, Adavere and Jaani Stage. The regression episodes in Raikküla, Rootsiküla and Paadla Stages correspond to minimum amount of clay in carbonates (Viiding, 1995).

The Silurian sequence in Estonia spreads on Hiiumaa, Saaremaa and Muhu islands and in the central part of Estonia. Thicknesses of the Silurian

Table 4. Stratigraphy and lithological characteristics of Estonian Ordovician and Silurian systems (Raukas & Teedumäe, 1997).

System	Series	Stage	Characteristic rocks
SILURIAN	Přidoli	Ohesaare	Clayey limestone, marlstone, dolomite, domerite
		Kaugatuma	Detritic limestone, marlstone, clayey limestone
	Ludlow	Kuressaare	Clayey limestone, marlstone, domerite
		Paadla	Detritic limestone, micritic limestone, dolomite, clayey limestone, marlstone, domerite
	Wenlock	Rootsiküla	Limestone, dolomite, domerite
		Jaagarahu	Detritic limestone, micritic limestone, clayey limestone, marlstone, dolomite
		Jaani	Clayey limestone, marlstone, calcareous clay, domerite, clayey dolomite
	Llandovery	Adavere	Marlstone, domerite, clay, detritic limestone, clayey limestone, dolomite
		Raikküla	Detritic limestone, aphanitic limestone, micritic limestone, dolomite, marlstone
		Juuru	Micritic limestone, detritic limestone, clayey limestone, marlstone, domerite
ORDOVICIAN	Upper Ordovician	Porkuni	Detritic limestone, sandy limestone, dolomite, marlstone
		Pirgu	Clayey limestone, marlstone, aphanitic limestone
		Vormsi	Clayey limestone, marlstone
		Nabala	Aphanitic and clayey limestone
		Rakvere	Aphanitic and microcrystalline limestone
		Oandu	clayey limestone, marlstone, detritic and micritic limestone
		Keila	Detritic limestone, clayey or microcrystalline limestone
		Haljala	Clayey limestone, marlstone, detritic limestone
		Kukruse	Limestone, clayey limestone, kukersite
	Middle Ordovician	Uhaku	Clayey limestone, marlstone with interlayers of kukersite
		Lasnamägi	Detritic limestone, dolomitized limestone
		Aseri	Oolitic limestone, detritic limestone
		Kunda	Detritic limestone, sandy limestone, dolomitized limestone, oolitic limestone
		Volkhov	Glauconitic limestone, microcrystalline limestone, Dolomitized limestone
	Lower Ordovician	Billingen	Glauconitic limestone
		Hunneberg	Glauconitic sandstone
		Varangu	Clayey siltstone, argillite
		Pakerort	Argillite, siltstone, sandstone, obolus sandstone

sediments are measured up to 436 m in Saaremaa (Viiding, 1995). The Silurian sequence consists mostly of carbonate rocks – limestones, dolomites and marlstones and is divided into ten regional Stages, representing all four global Series (Table 4).

The Silurian and Ordovician stratigraphic record in Estonia contains numerous volcanic ash layers whose composition is dominated by the mineral assemblage of illite – smectite – K-feldspar - kaolinite. The contents of these minerals may vary between individual bentonite layers. Clay-rich interbeds of volcanic origin in Palaeozoic sedimentary sections are commonly called K-bentonites. K-bentonite is a product of chemical weathering of volcanic ash, where an apposition from terrigenous material is found. The seawater has turned into the strongest factor in the process of chemical weathering (Jürgenson, 1958). The main mineralogical characteristics of K-bentonites are: presence of illite-smectite and kaolinite; low content or absence of quartz; low content or absence of carbonate minerals; a high content of authigenic K-feldspar (low-sanidine); presence of visually detectable biotite flakes, commonly absence of chlorite; presence of zircon and apatite. The thicknesses of K-bentonites vary from some mm up to 10 cm, rarely up to 20-30 cm (Kiipli et al., 1997).

The interlayers of K-bentonite have been recorded in Jaani, Adavere and Juuru Stages of Silurian System (Jürgenson, 1964; Kiipli et al., 2001; Kiipli & Kallaste, 1996; Kiipli, 1998). Only the Adavere Stage itself contains 27 volcanic beds, the most prominent of which is the “O” K-bentonite bed with thickness up to 25 cm. Ordovician K-bentonite beds in Estonia are present only in the Upper Ordovician Series: Kukruse, Haljala, Keila and Pirgu Stages (Bergström et al., 1995; Kiipli et al., 2004). Most outstanding of them is the Kinnekulle bentonite bed with thickness up to 70 cm in Saaremaa Island.

THE SILURIAN-ORDOVICIAN AQUIFER SYSTEM

The Silurian-Ordovician (S-O) aquifer system consists of diverse limestones and dolomites with clayey interlayers and it is an important and often the only source of drinking water in central and western Estonia and on the islands of the West-Estonian Archipelago.

The upper part of the S-O aquifer system with a thickness of 30 m is extremely cavernous, with numerous cracks and fissures. Karst cavities form some half-a-metre-deep canals trending in the direction of bedrock fissures. Caverns are especially abundant in dolomites and dolomitized limestones. Close to the ground surface, bigger karst cavities, a couple of metres deep and some twenty or thirty meters long, occur in some places. Water in the fissure systems and karst cavities of the carbonate bedrock flows relatively fast. Groundwater recharges from the Quaternary deposits in the outcrop area and therefore can easily become polluted in the areas with a thin Quaternary cover. Silurian and Ordovician carbonate rocks have fragmentary water conducting zones with parallel lamination and an abundance of fissures. In these 1-2 m thick zones groundwater flows in a lateral direction. Zones with high lateral conductivity are separated from each other by 5-10 m thick layers - aquitards in which groundwater flows mainly in vertical fissures (Perens & Vallner, 1997; Heinsalu, 1995). Fractures and karst are most common near the ground surface, downwards the fracturing of carbonate rocks decreases. In the topmost 20 m of carbonate aquifer the lateral conductivity is 10-50 m/d, 5-8 m/d at a depth of 20-50 m and only 1-2 m/d at a depth of 50-100 m (Perens & Vallner, 1997; Heinsalu, 1995). The hydraulic conductivity of rocks decreases with depth and S-O aquifer system transforms into S-O aquitard.

The aquifer system has a characteristic $\text{HCO}_3\text{-Ca-Mg}$ and $\text{HCO}_3\text{-Mg-Ca}$ water type with TDS mainly below 0.6 g/L in its upper 30-50 m thick portion. In coastal areas and greater depths the content of Cl^- and Na^+ in groundwater increases and $\text{HCO}_3\text{-Cl-Na-Mg-Ca}$ type water with TDS between 0.3-1.5 g/L is widespread (Perens et al., 2001).

FLUORIDE CONTENTS IN SILURIAN-ORDOVICIAN AQUIFER SYSTEM

All aquifer systems found in the sedimentary cover are used in drinking and domestic water supply in Estonia. Estonian hydrostratigraphical cross-section begins with Quaternary aquifer system, followed by Upper-Devonian, Middle-Devonian, Middle-Lower-Devonian, Silurian-Ordovician, Ordovician-Cambrian and Cambrian-Vendian aquifer system, that are isolated from each other by impervious beds. Aquifer systems differ from each other in distribution, bedding conditions, hydraulic parameters and chemical composition.

Four hundred thirty (430) groundwater samples were collected from Silurian-Ordovician aquifer system, whence 48.2 % does not comply with Estonian as well as EU and WHO requirements in respect of F^- (1.5 mg/L). On the other hand, 19.4 % of the drinking water samples contained fluoride less than 0.5 mg/L. Consumption of drinking water with such low fluoride content is insufficient to prevent caries.

Figure 1 shows the spatial distribution of the investigated drinking water wells, where fluoride levels range from 0.01 to 7.2 mg/L. An analysis of the regional distribution of fluoride concentrations shows a great variation. Northern Estonia as well as eastern Estonia are characterised predominantly with low fluoride content in drinking water, with exception only in Kernu district (3.6 mg/L). Fluoride concentrations in Järva and Jõgeva County reach up to 2.24 mg/L and 5.0 mg/L, respectively. Fluoride contents exceed the limit value set for drinking water also in Viljandi ($F=0.32-2.46$ mg/L) and Rapla (0.1-2.75 mg/L) County. It is possible to delimit an area with anomalously high F^- content, which primarily involves Pärnu and Lääne Counties (Fig. 1). Almost half of the Pärnu districts have F-rich groundwaters, what is a serious problem for local administrations and water-companies. High concentrations of F^- are determined in wells of Tõstamaa, Audru, Koogna, Halinga, Are, Kaisma, Tori and Paikuse districts. Highest F^- values (up to 7.2 mg/L) have been measured in Audru, Tõstamaa and Pärnu-Jaagupi. Majority of the wells in western Estonia have F^- concentrations over the limit value; the highest values are detected in Hanila, Taebala and Lihula districts, where fluoride concentrations

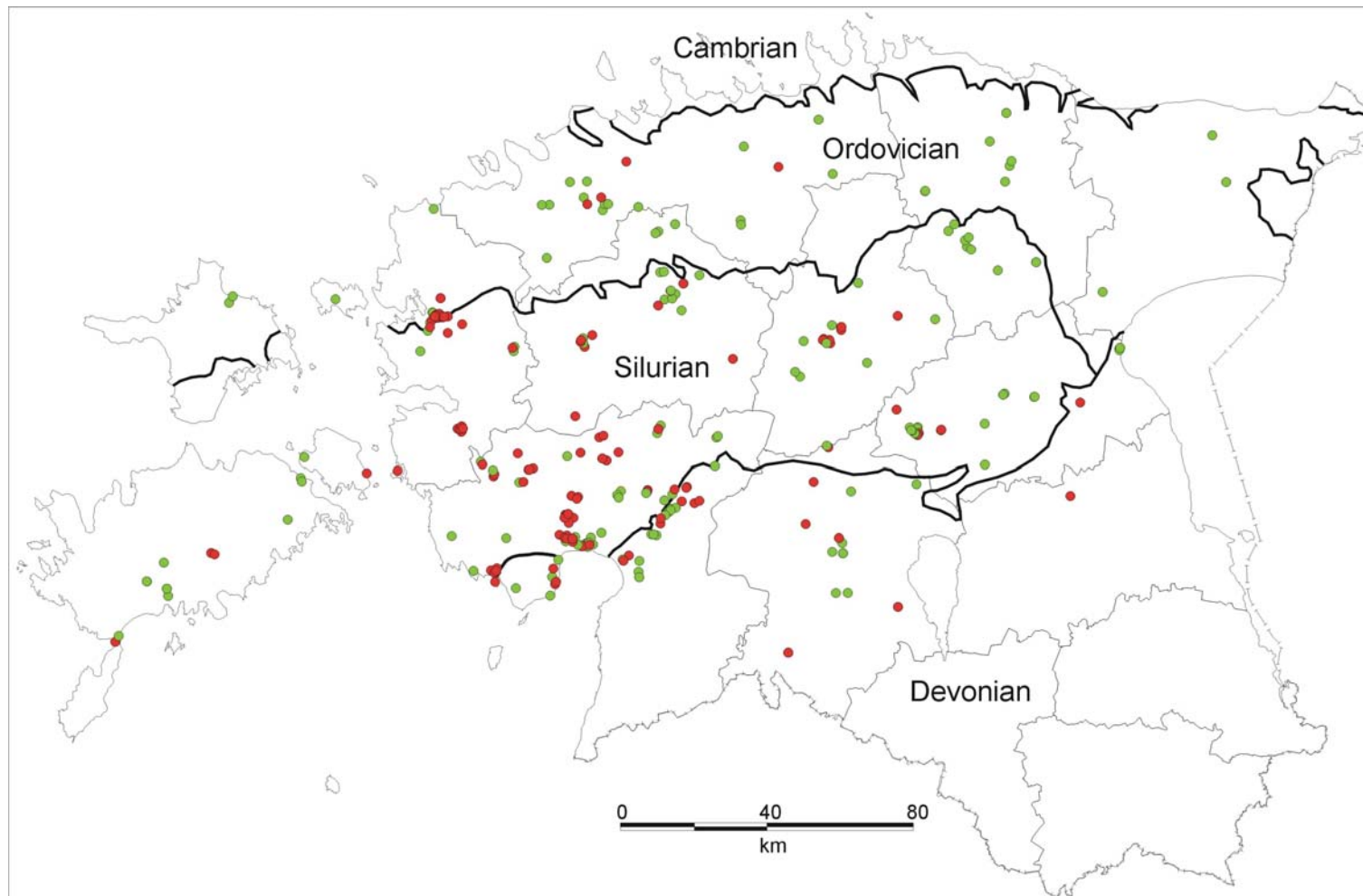


Fig. 1. Groundwater sampling sites are marked on the background of schematic geological map. Fluoride concentrations over 1.5 mg/L are marked with red spots.

reach up to 5.7 mg/L. Therefore, high F^- concentrations of natural origin are common in western Estonia, where drinking and domestic water originates from aquifers associated with Ordovician and Silurian carbonate rocks.

Vertical distribution of fluoride in groundwater

High fluoride concentrations can be achieved in groundwaters that have long residence times in the host aquifers. Shallow groundwaters usually have low concentrations of fluoride because they represent recently infiltrated rainwater. Deeper (older) groundwaters with long residence time contain therefore most likely high concentrations of fluoride. This is in accordance with analytical results, which show that the highest concentrations of fluoride (up to 7.2 mg/L) are analysed in wells having a depth of 150-200 m (Fig. 2). The lowest F^- values are detected in shallow wells (<30 m) since water flows relatively fast and is strongly affected by infiltration of rainwater. Such

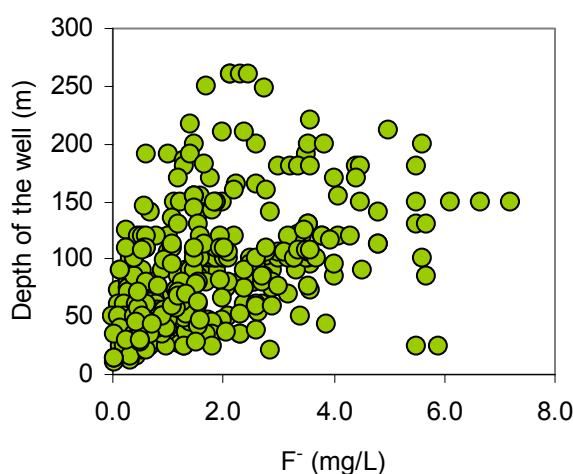


Fig. 2. Bivariate plot of F^- versus well depth in Silurian-Ordovician aquifer system.

tendency is explained by geological characteristics of Silurian-Ordovician aquifer system. The upper part of the aquifer system with a thickness of 30 m is extremely karstified and fissured. Karst phenomena and the frequency of fissures decrease downwards, as the total transmissivity of carbonate rocks.

Generally, the amount of dissolved chemical substances in groundwater increase with depth, because of the deceleration of water exchange and a longer contact time with host rocks. In reality, there also exist deviations from the general tendencies that can be explained with the differences in the constructions of wells, their open intervals and also in the type and structure of the opened rocks.

Vingisaar et al. (1981) have been studied the chemical composition of Estonian Silurian and Ordovician carbonate rocks. Correlation of fluoride content with lithology, particularly with siliciclastic clay and dolomite content is shown in the lithological diagram (Fig. 3). Fluoride content is remarkable higher in dolomites compared to limestones. If clay content of the rock increases, it is followed by the increase on F^- content, therefore the maximum amount of fluoride can be found in clayey dolomites. The average F^- contents in pure limestone, dolomite and clayey dolomite are 650, 2000 and 2500 g/t, respectively (Fig. 3). Comparing the geological cross-sections and water chemistry of boreholes the same trend can be followed in western Estonia.

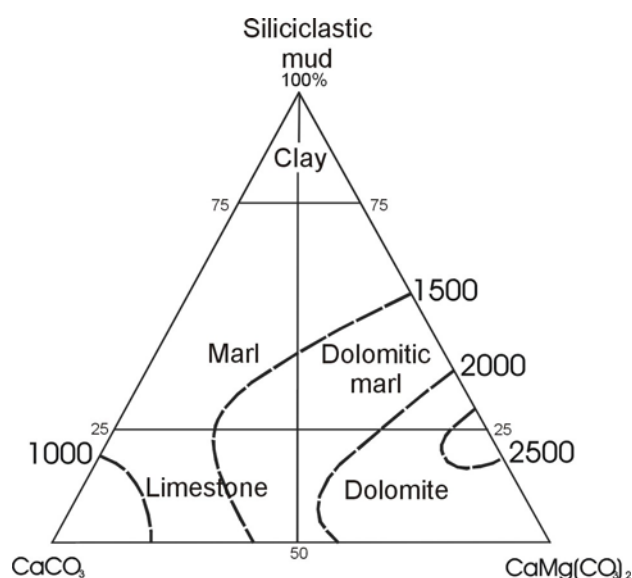


Fig. 3. The lithological diagram of carbonate rocks. Isolines reflect the fluoride content (g/t) in the rocks (Vingissaar et al., 1965; 1981).

The boreholes (Figs. 4, 5 and 6) with the same depth but different lithological composition enable to make assumptions on the variation of groundwater fluoride content. For example, fluoride content in Audru borehole (well no. 6189) is 6.1 mg/L, where dominating rock types are dolomite and clayey limestone. Well with the same depth in Märjamaa (well no. 8718) exhibits fluoride content almost three times lower than in Audru, because the dominating rock types there are mainly limestone and clayey limestone. Similar principle exists in 170 m (well no. 6671 and 9737) and in 105 m (well no. 6671 and 9737) deep boreholes. Low fluoride concentrations can be found in boreholes that open limestones and high contents correlate with the presence of marl and dolomite interlayers in limestones. Even shallow wells (well no. 7923, 6154 and 6586) drilled into rocks of Adavere and Jaagarahu Stage could show high F contents. As a rule, there occurs predominantly intensive water exchange in shallow wells and therefore low concentration of chemical components. However, in places thick Quaternary cover prevents from infiltration of rainwater and consequent dilution of groundwater, which

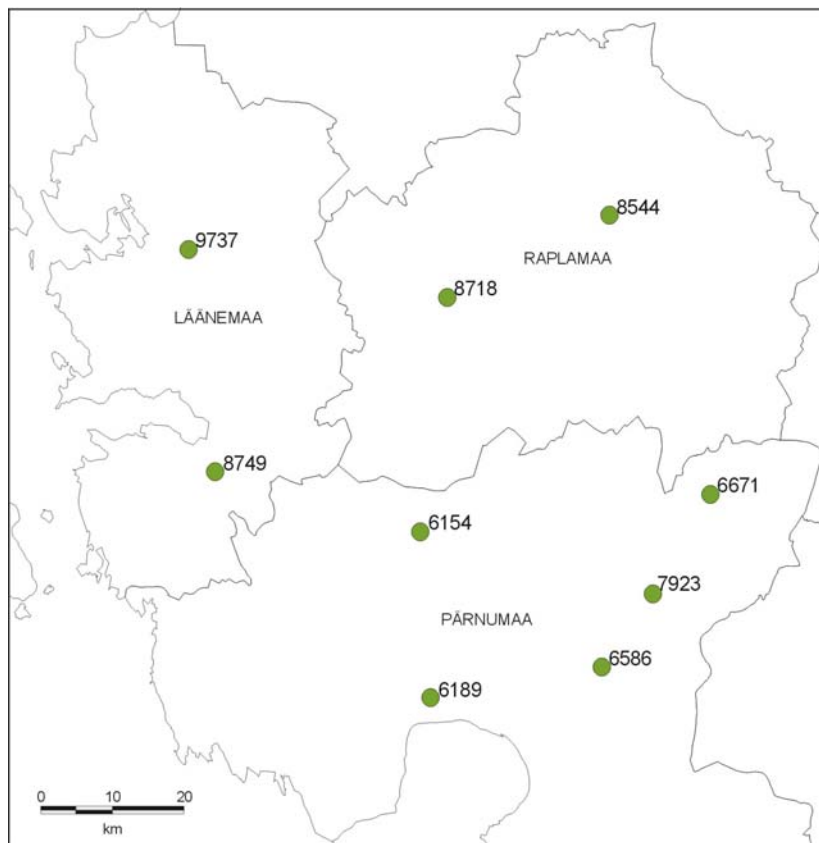


Fig. 4. Location of the boreholes (Figs. 5 and 6).

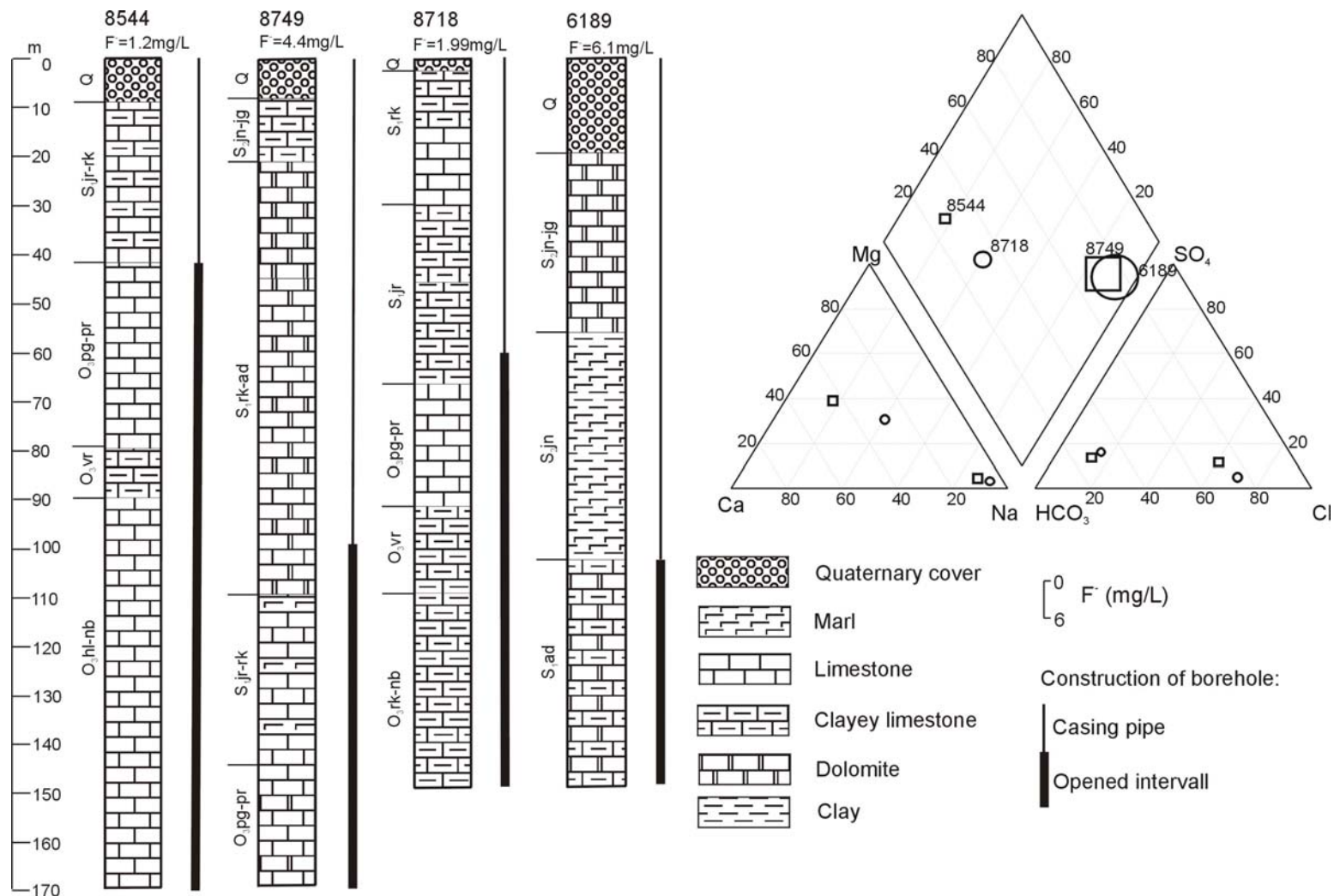


Fig. 5. Geological cross-sections of boreholes (cadastre number 8544, 8749, 8718 and 6189) and groundwater chemistry.

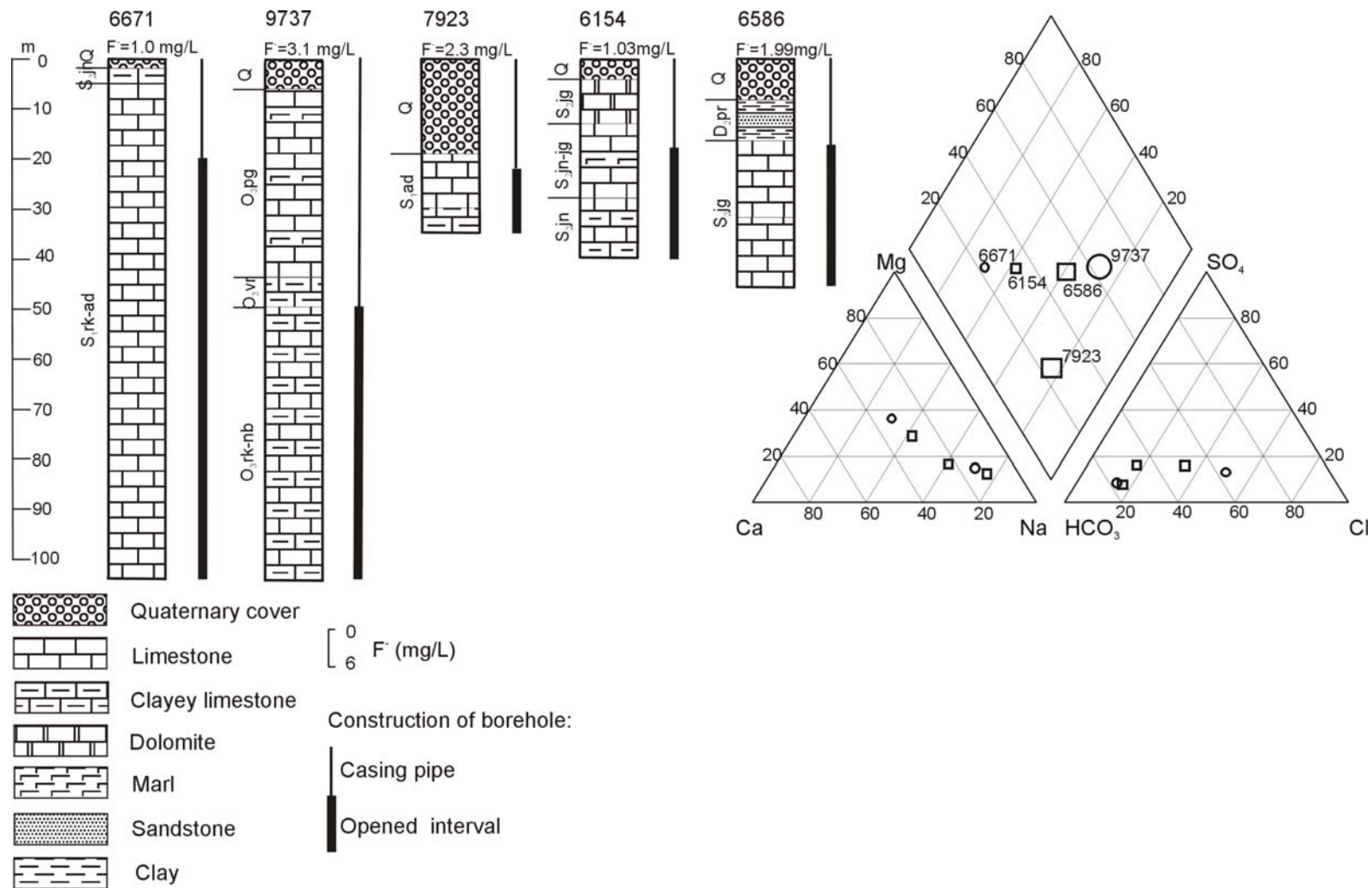


Fig. 6. Geological cross-sections of boreholes (cadastre number 6671, 9737, 6154, 6586) and groundwater chemistry.

could lead to F-rich groundwaters in small depths (well no. 7923 and 6586). Therefore, the intensity of water exchange acts as an important factor regulating the content of F in groundwater.

The F contents of the water are in good correlation with the length of the open intervals of boreholes (Figs. 5 and 6). Wells with low fluoride concentration have twice longer open intervals than the wells with the same depth but high fluoride contents. The long open interval of the bored well enables the mixing and dilution of waters with different salinity and fluoride content. Therefore, the higher F^- contents are related to deep wells with short screens.

Since, in the open interval of bored wells several rock types are often presented, no apparent correlation between F content and host rocks exists. The correlation between groundwater and host rock chemistry might be stronger, if the open intervals of wells are shorter and only one rock type is exposed.

Hydrochemistry of fluoride

The chemical composition of groundwater is affected by concurrence of several factors. Downwards the fissures of carbonate rocks and the water exchange rate decrease. Groundwater that has long residence time in host rocks reflects the chemical composition of rocks, geochemistry of aquifer and hydrodynamics.

According to Lahermo et al. (1991) groundwaters with high F^- contents are generally HCO_3^-Na -type waters, particularly poor in Ca^{2+} . Several authors have shown that in waters with high F concentrations, the amount of F is proportional to the HCO_3^- concentration and the pH (Handa, 1975; Saxena & Ahmed, 2001). Same trend can be observed in Silurian-Ordovician aquifer system, where elevated F concentrations are mostly associated with pH values over 7.6 (Fig. 7). The pH values from 7.6 to 8.6 and high HCO_3^- contents (350-450 mg/L) are favourable for fluoride dissolution.

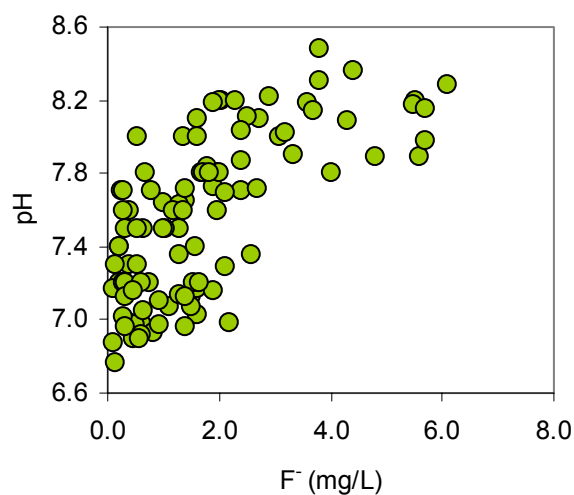


Fig. 7. The pH values plotted against fluoride content in Silurian-Ordovician aquifer system.

Groundwater in Silurian-Ordovician aquifer system is mainly HCO_3 -Ca-Mg-type (Perens & Vallner, 1997) and owing to the high Ca^{2+} contents, quite low amounts of F^- may be mobilised. The highest F^- concentrations exist in wells that produce the water with low Ca^{2+} content (Fig. 8).

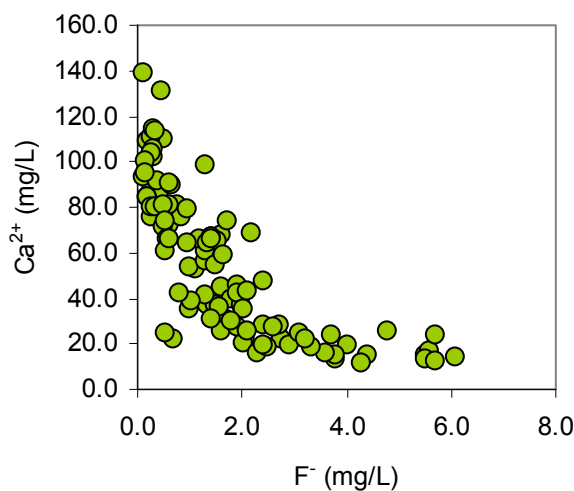


Fig. 8. The Ca^{2+} values plotted against fluoride content in Silurian-Ordovician aquifer system.

The dissolution and deposition of fluoride in groundwater is controlled by chemical type of water, which varies considerably in the study area. Therefore, major ion chemistry of sampled groundwater is examined as a

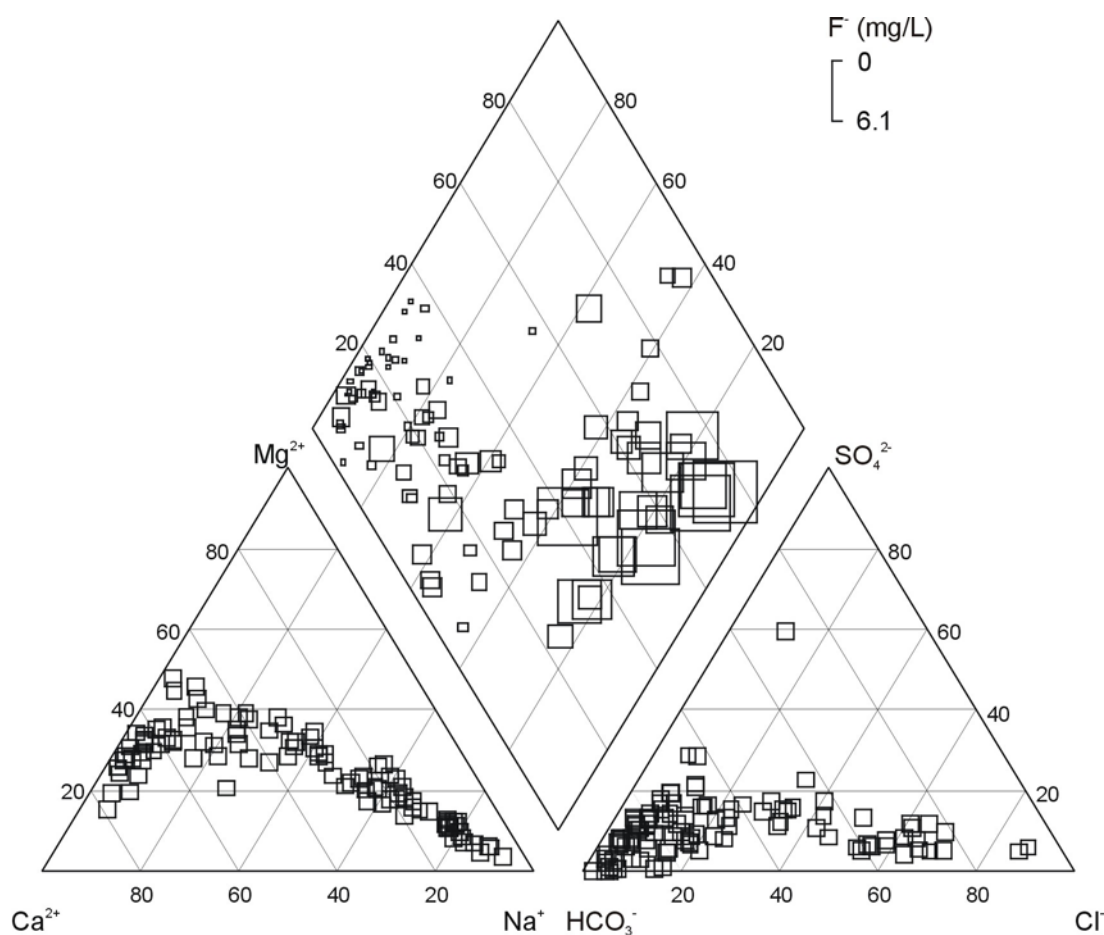


Fig. 9. Piper diagram reflecting the chemical type of groundwater and proportional content of F^- .

whole using the Piper diagram (Fig. 9). In order to construct the Piper diagram main cations (Na^+ , Ca^{2+} , Mg^{2+}) and anions (Cl^- , SO_4^{2-} , HCO_3^-) are counted. In Piper diagram the concentrations of main ions are expressed as % meq/L and it shows direct relation between proportional fluoride content and the chemical type of groundwater. High fluoride concentrations are associated with HCO_3^- -Cl-Na or HCO_3^- -Cl- SO_4 -Na type waters. The lowest fluoride values can be found in the HCO_3^- - SO_4 -Ca-Mg type waters, where F concentrations are below 0.5 mg/L. Groundwater with HCO_3^- -Na chemical type has always very high

fluoride content. As the concentration of calcium in groundwater increases the concentration of fluorine decreases.

Relation between fluoride content and groundwater chemical type is expressed on hydrogeochemical maps of Silurian and Ordovician aquifers, where proportional fluoride concentrations are marked (Figs. 10 and 11). Cations and anions that have equivalent concentrations higher than 20 % are plotted on the hydrogeochemical map. The groundwater chemistry varies remarkably in different parts of the study area. In the Silurian aquifer HCO_3^- Ca-Mg type groundwater is mainly spread in Rapla County, where the concentration of F^- in groundwater is low. In the Lihula district high F^- values (1.6-4.4 mg/L) are associated with Cl- HCO_3 -Na-Mg or Cl- HCO_3 -Na type groundwaters, where favourable hydrochemical conditions for dissolved F^- in groundwater exist. Practically three hydrochemical types can be distinguished in Pärnu County (F^- =0.01-7.2 mg/L): Cl- HCO_3 -Na-Mg-, Cl- HCO_3 -Na type groundwater, HCO_3 -Mg-Ca type and HCO_3 -Cl-Na-Mg-Ca-, HCO_3 -Cl-Na type water. As the hydrochemical types gradually change to the bicarbonate calcium magnesium (HCO_3 -Ca-Mg) or bicarbonate sodium magnesium calcium (HCO_3 -Na-Mg-Ca) waters, the concentration of F^- falls. Therefore Viljandi, Jõgeva and Lääne-Viru County are mainly represented by F^- poor groundwater. Vertical zonation is also characteristic to the Silurian aquifer that is expressed by the downward increase of Cl^- and Na^+ content and mineralization of water. Therefore, some derivations may occur between groundwater chemical type and F^- content.

In the Ordovician aquifer changes in fluoride concentration are also followed by changes in general water chemistry. In chloride bicarbonate sodium magnesium (Cl- HCO_3 -Na-Mg), chloride bicarbonate sodium (Cl- HCO_3 -Na), bicarbonate chloride sodium magnesium calcium (HCO_3 -Cl-Na-Mg-Ca) and bicarbonate chloride sodium (HCO_3 -Cl-Na) waters the concentration of F^- in groundwater is high. As the hydrochemical types change to the HCO_3 -Mg-Ca, HCO_3 -Na-Mg-Ca and HCO_3 -Ca-Mg waters, the concentration of F^- falls. Accordingly, high values of fluoride can be attributed to Lääne County, where Cl- HCO_3 -Na-Mg and Cl- HCO_3 -Na type waters dominate. Low levels of fluoride

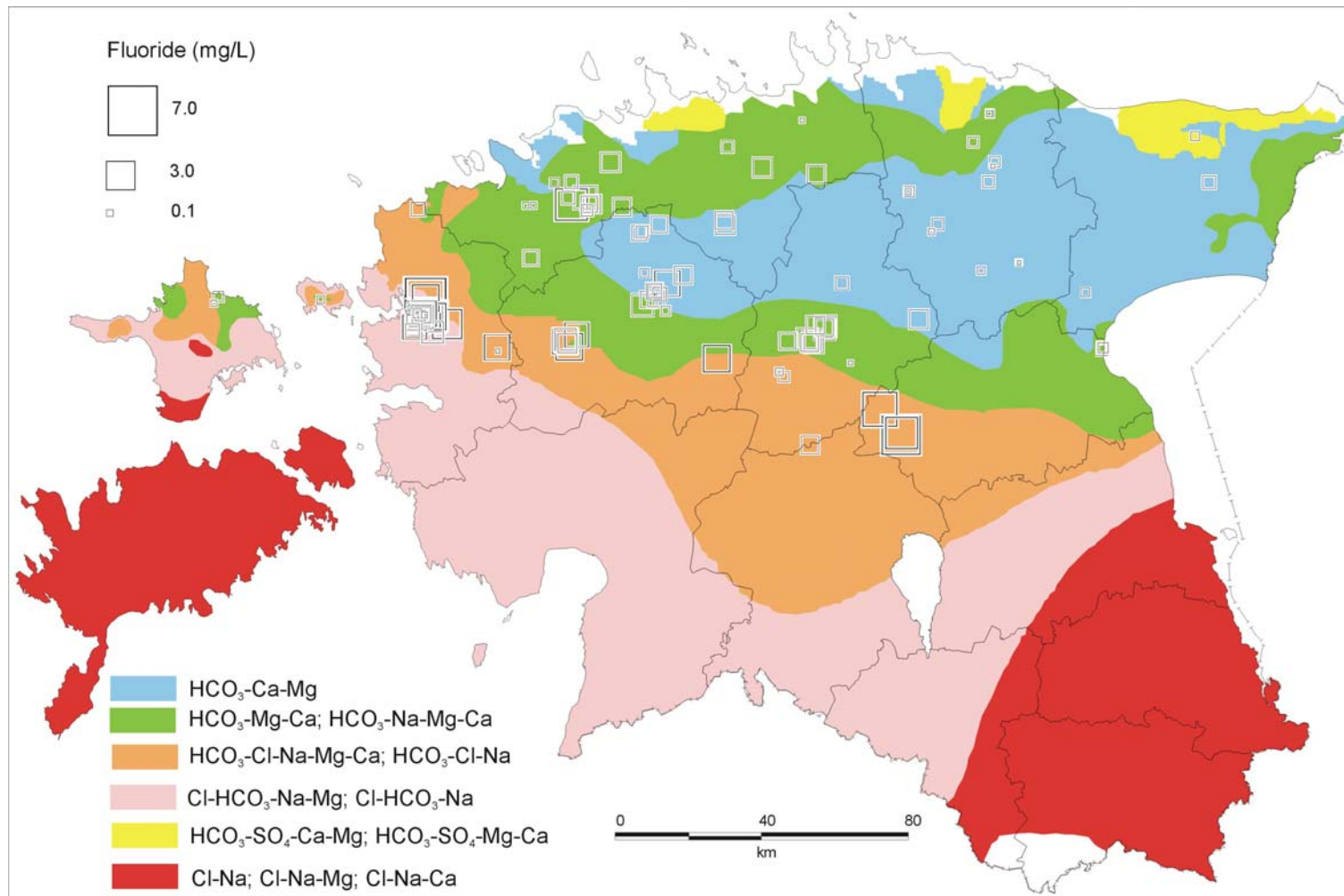


Fig. 10. The concentrations of fluoride on the hydrogeochemical map of Ordovician aquifer system. Groundwater chemical types after Perens et al., 2001.

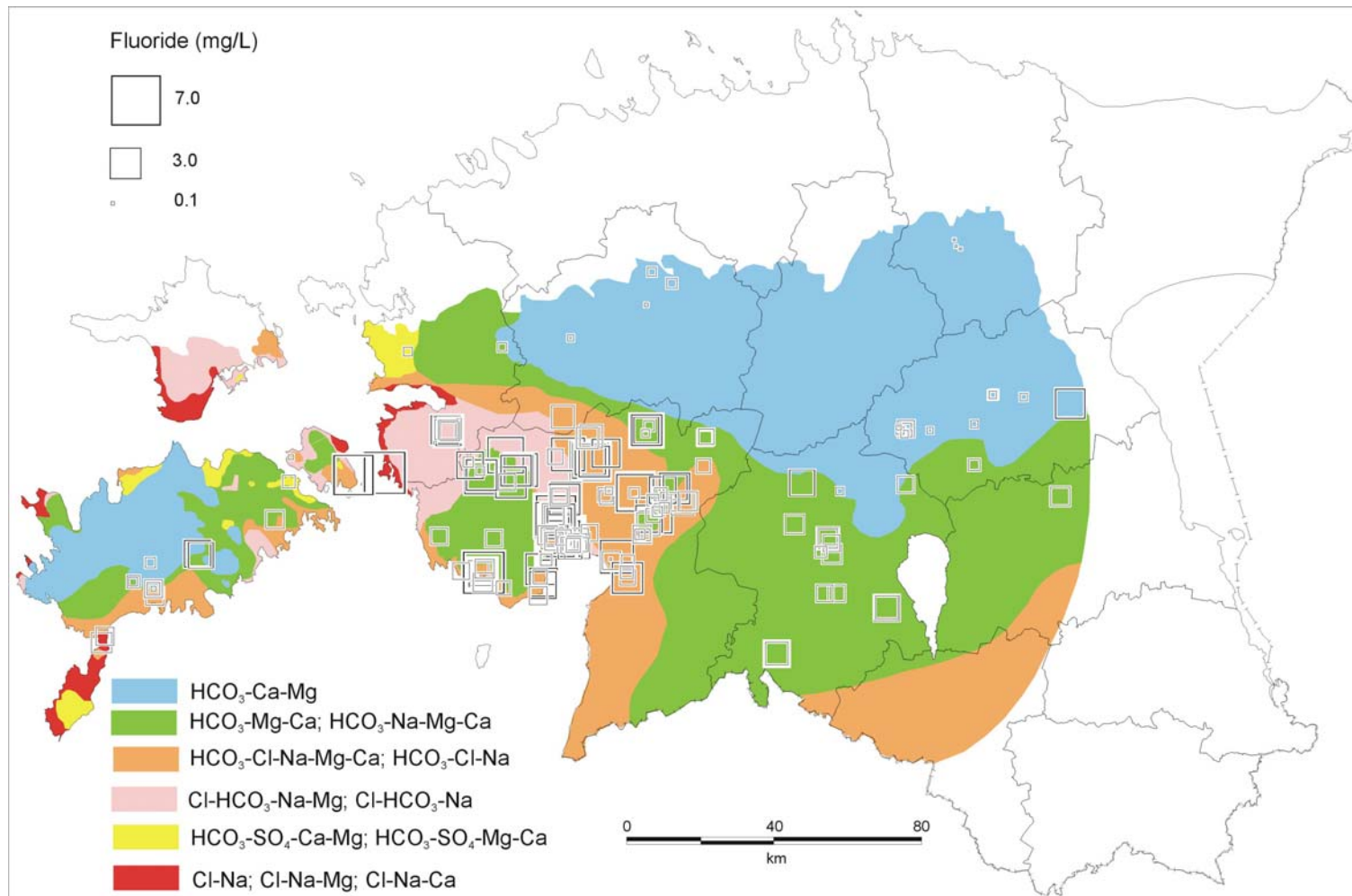


Fig. 11. The concentrations of fluoride on the hydrogeochemical map of Silurian aquifer system. Chemical types of groundwater after Perens et al., 2001.

in the groundwater are determined in Ida-Viru and Lääne-Viru County, elsewhere in the Country F^- concentrations remain below the permitted limit value (1.5 mg/L).

Deviations from the above mentioned general tendencies can be explained by the differences in the depth and construction of wells and also in the type and structure of the opened rocks. Therefore, the occurrence of wells with different groundwater chemical type close to each other is possible.

Fluoride concentrations are relatively independent of the other water-soluble components, but have noteworthy correlation with boron (B). Highest F values coincide with anomalously high B concentrations and are analysed from Silurian-Ordovician aquifer system (Fig. 12).

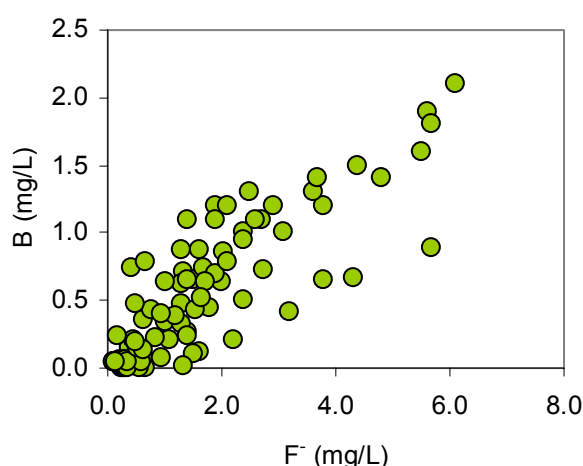


Fig. 12. The concentrations of B plotted against fluoride concentration.

Since ocean water contains abundantly boron (4.4 mg/L) beside fluorides, the correlation between these two elements in groundwater is anticipated. Boron concentrations range from 0.016 to 2.1 mg/L, with a spatial distribution roughly similar to that of fluoride. The upper officially permissible B concentration in drinking water in Estonia is 1.0 mg/L (Joogivee, 2001). This value is often exceeded in western Estonia where boron-rich groundwaters may cause different health hazards - mild gastrointestinal irritation (Saava, 1998).

POSSIBLE SOURCE OF FLUORIDE-RICH GROUNDWATER

Various studies of volcanic emissions have shown that volcanic gases and tephra are rich in F (Barclay et al., 1996; Horn & Schmincke, 2000; Weinstein & Cook, 2005). In 1980 Óskarsson demonstrated that the tephra from the 1970 eruption of Hekla (Iceland) contained up to 2000 mg F/kg. In leaching studies on the Hekla ash (2 g tephra in 100 mL water), after 15 min of shaking, 1000 µg F was released per g of tephra. From fresh samples of Guatemalan volcanoes an average of 14–29 µg F/g was leached into water during the first step of a three-step successive leaching study (Smith et al., 1982). A study of tephra from the Ruapehu volcano in New Zealand showed that the F concentration was as high as 67.9 mg/kg after the 24 hr of leaching with water (Cronin et al., 2003). Ion probe analyses of matrix glasses and melt inclusions of different minerals in tephra of the tenth century A.D. eruption of the Baitoushan volcano (China/North Korea) revealed F contents of 2430–3990 mg/kg and 2630–6200 mg/kg, respectively (Horn & Schmincke, 2000).

It might be supposed, that the volcanic material in carbonate rocks is one of the most probable source of high F⁻ in groundwater. The interlayers of K-bentonite have been recorded in Adavere and Jaani Stages of Silurian system in the study area. The layers of volcanic ashes have been also found in Kukruse, Haljala, Keila and Pirgu Stages from the cross-sections of Ordovician system.

Fluoride concentration in 25 rock samples was studied in ten drillcores of Estonia – Paatsalu, Ikla, Varbla, Pärnu, Laitse, Valga, Laeva, Pürksi, Ingliste and Oru (Figs. 13, 14 and 15). The results of the chemical analyses and leaching experiments are presented in the Table 5. The content of F leached into water from carbonate rocks (limestone, dolomite) varies between 4–10 mg/kg and is remarkably higher in clayey dolomite (26.8 mg/kg). The total F content of studied carbonate rocks is mostly 100–400 mg/kg, being highest in clayey dolomite (1100 mg/kg). The concentration of F in seawater (salinity 35‰) is 1.3–1.4 mg/L (Skinner & Turekian, 1973). Calcium carbonate

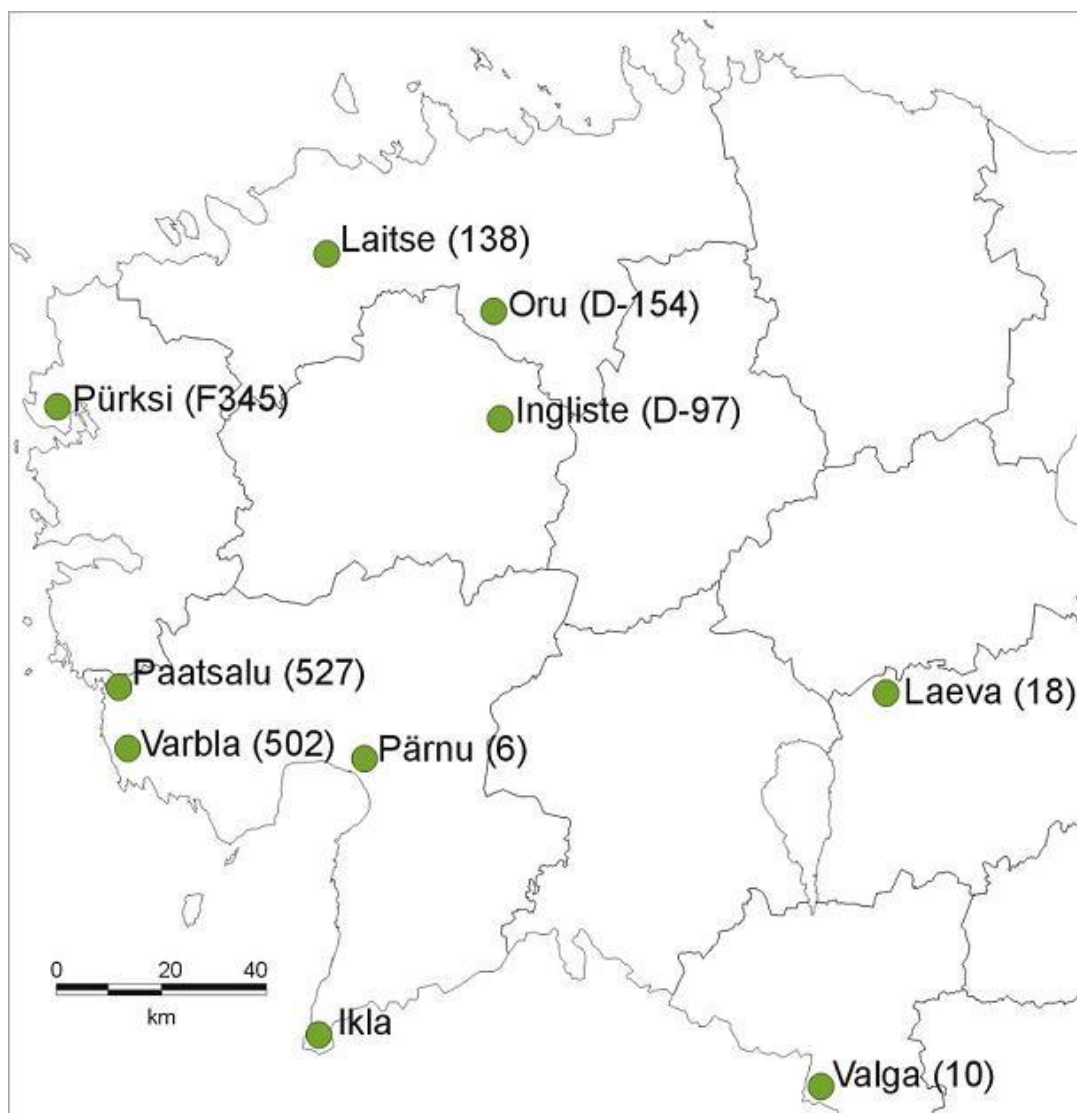


Fig. 13. Locations of studied cross-sections (Figs. 14 and 15).

precipitation dominates the removal of dissolved fluoride from seawater and remarkable amounts of fluorides will be incorporated into clay minerals (Carpenter, 1969). Therefore, the dissolution of fluorides from carbonate rocks is one probable source of fluoride-rich groundwater.

K-bentonites are rich in total F – 1400-4500 mg/kg and several times higher amounts of F (25-52 mg/kg) were leached into water compared to carbonate rocks (Table 5). During the second extraction (0.5 M Tiron) 33-191 mg/kg of F associated with oxidic Fe and Al sites was solubilised from K-bentonites.

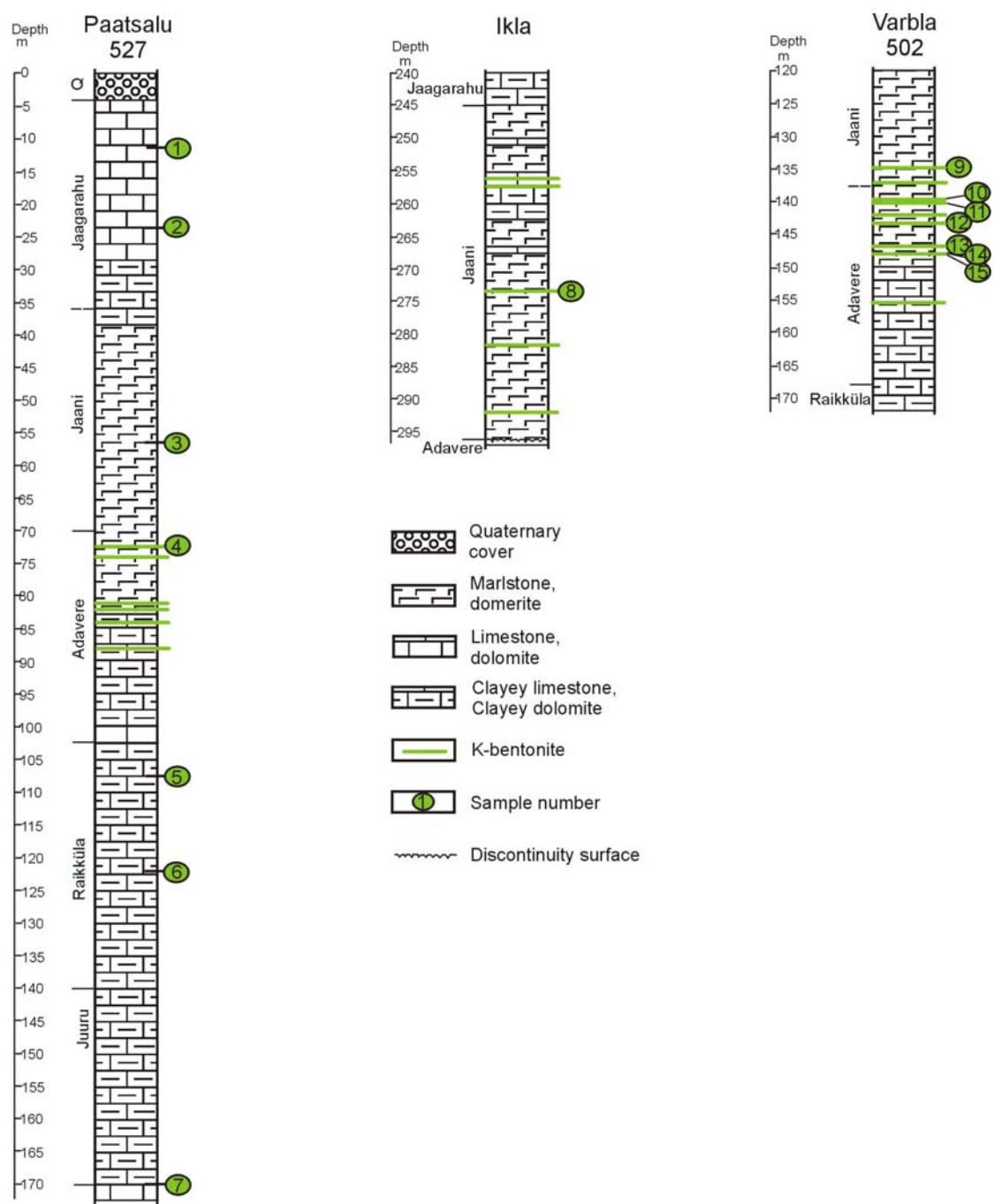


Fig. 14. The lithology of studied Silurian core sections and locations of sampled carbonates and K-bentonites. Paatsalu core section after Nestor, 1997; Ikla core section after Nestor & Nestor, 2002; Varbla core section after Nestor, 1994.

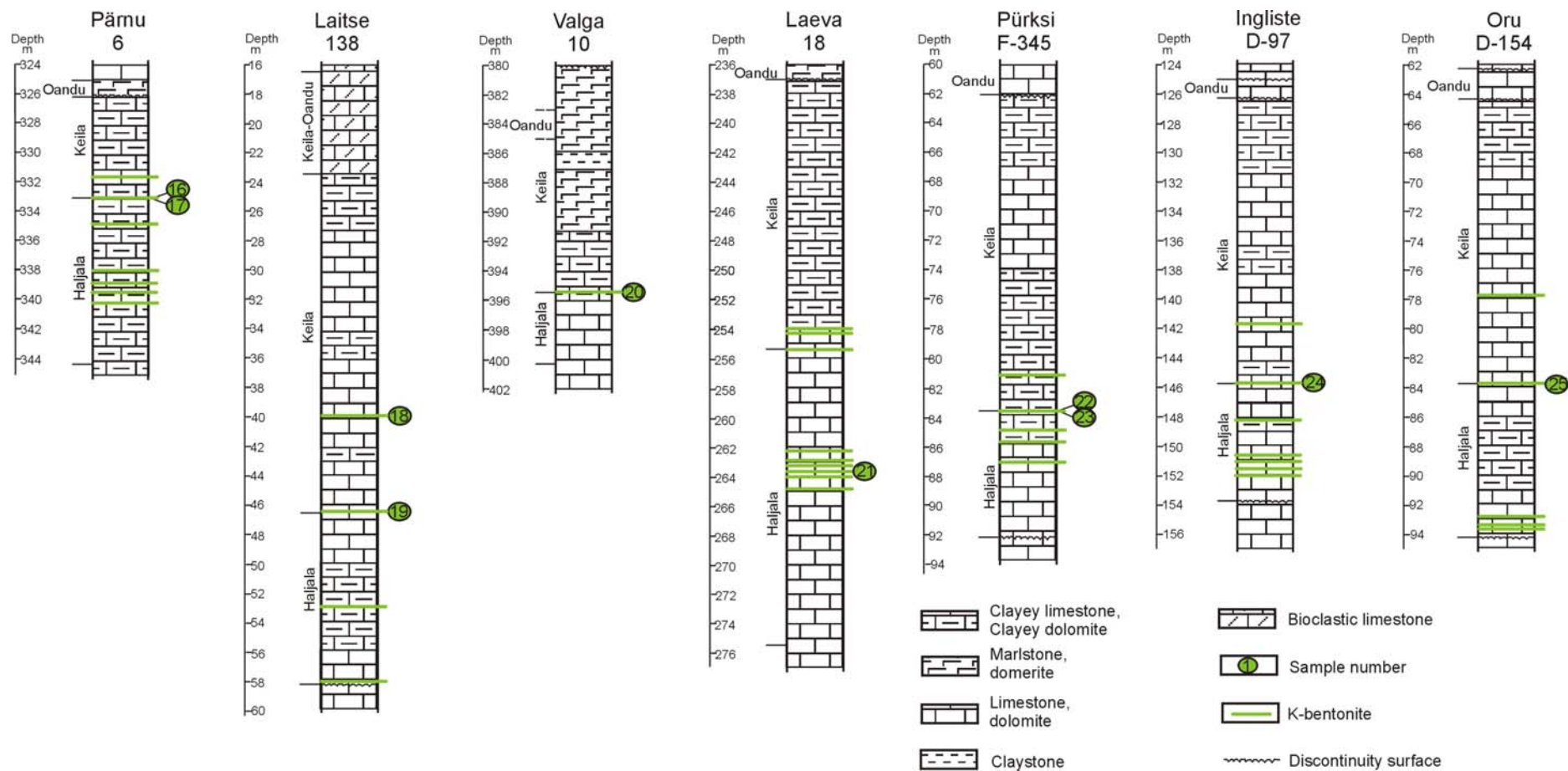


Fig. 15. The lithology of studied Ordovician core sections and locations of sampled K-bentonites. The lithology of studied core sections by Leho Ainsaar (unpublished data).

Table 5. The results of the chemical analyses. The total F content of the rocks (F_{Tot}) and F concentrations leached into water (F_{Water}) and Tiron solution (F_{Tiron}). Numbering of samples as shown in Figs. 14 and 15.

SAMPLE NO.	CORE	DEPTH (m)	STAGE	ROCK	F_{Tot} (mg/kg)	F_{Water} (mg/kg)	F_{Tiron} (mg/kg)
1	Paatsalu 527	11.50	Jaagarahu	dolomite	300	3.86	
2	Paatsalu 527	23.50	Jaagarahu	dolomite	100	4.26	
3	Paatsalu 527	56.30	Jaani	clayey dolomite	1100	26.80	
4	Paatsalu 527	72.50	Adavere	K-bentonite	4400	39.30	98.7
5	Paatsalu 527	107.50	Raikküla	limestone	200	5.19	
6	Paatsalu 527	122.10	Raikküla	limestone	400	4.31	
7	Paatsalu 527	167.30	Juuru	limestone	300	9.79	
8	Ikla	273.50	Jaani	K-bentonite	1900	27.00	66.3
9	Varbla 502	134.90	Adavere	K-bentonite	1400	25.10	19.5
10	Varbla 502	139.85	Adavere	K-bentonite	4000	50.40	71.0
11	Varbla 502	140.25	Adavere	K-bentonite	2800	31.80	63.8
12	Varbla 502	143.25	Adavere	K-bentonite	2800	30.30	126.2
13	Varbla 502	146.90	Adavere	K-bentonite	3200	40.80	59.9
14	Varbla 502	148.00	Adavere	K-bentonite	3200	26.90	56.8
15	Varbla 502	148.20	Adavere	K-bentonite	4000	28.30	97.7
16	Pärnu 6	333.30	Keila	K-bentonite	4500	41.50	90.5
17	Pärnu 6	333.45	Keila	K-bentonite	4400	35.40	92.2
18	Laitse 138	39.90	Keila	K-bentonite	1700	24.90	137.1
19	Laitse 138	46.30	Keila	K-bentonite	4100	29.40	80.3
20	Valga 10	395.20	Keila	K-bentonite	1600	35.90	47.1
21	Laeva 18	263.50	Haljala	K-bentonite	2900	41.40	71.7
22	Pürksi F-345	83.30	Keila	K-bentonite	3400	34.20	33.3
23	Pürksi F-345	83.50	Keila	K-bentonite	3100	38.00	52.6
24	Ingliste D-97	145.80	Keila	K-bentonite	3800	34.20	191.5
25	Oru D-154	84.00	Keila	K-bentonite	4200	51.90	48.7

Thus, the total F contents and the results of the leaching experiments of the current study coincide well with previously published results. The high F concentrations are well preserved in altered volcanic ash beds and contribute to the formation of F-rich groundwater in the Silurian-Ordovician aquifer system.

CONCLUSIONS

Fluorides have widely been in the focus of public and scientific interest because of their important physiological role in the health of a man. Their beneficial effect in decreasing the incidence of dental caries is well known while the toxic effects of fluorides, when occurring in abundances in drinking water, cause increased worries. The fluoride concentrations in the wells analyzed, were found to deviate from the safe content range (0.8-1.2 mg/L) recommended by WHO as well as limit values set by EU and Estonian drinking water standards.

Silurian-Ordovician aquifer system is problematic because it does not meet the requirements for fluoride concentration and therefore is unqualified as a drinking water. The excess of fluoride in groundwater is of special concern in western Estonia because S-O aquifer system is often the only drinking water source in this region. It is possible to delimit an area with anomalously high F^- content (up to 7.2 mg/L) which primarily involves Pärnu and Lääne Counties. Health problems arising from excess fluoride intake (dental fluorosis) must be studied in this region.

The content of fluoride in Silurian-Ordovician aquifer system is related with the depth of the wells. High fluoride concentrations are associated with deep groundwaters that have long residence time in the host rock. Some deviations may also exist that can be explained with the differences in the constructions of wells, their open intervals and also in the type and structure of the opened rocks. The long open interval of the bored well enables the mixing and dilution of waters with different salinity and fluoride content. Therefore, the higher F^- contents are related to deep wells with short screens.

The fluoride dissolution and deposition in groundwater is controlled by chemical type of water, which varies considerably in the study area. High fluoride concentrations are associated with HCO_3 -Cl-Na or HCO_3 -Cl- SO_4 -Na type waters. Groundwater characterized by HCO_3 -Na type has always very

high fluorine content. As the concentration of calcium in groundwater increases the concentration of fluorine decreases.

Probable geological source of fluoride is the leaching of clayey carbonate rocks and clay-rich K-bentonite beds. The interlayers of K-bentonite have been recorded in Adavere and Jaani Stages of Silurian system in the study area. The layers of volcanic ashes have been also found in Kukruse, Haljala, Keila and Pirgu Stages from the cross-sections of Ordovician system.

Because of the toxicity of fluoride it would be necessary to undertake a systematic and detailed study of the relationship between water chemistry and the host lithology.

Groundwater abstracted from wells is directly, without fluoride removal, distributed between consumers. Because of that the location and the depth of new water supply wells must be selected very carefully to avoid need for later construction of expensive raw water treatment facilities and increase in the price of water. The wells that are operating at the moment should be carefully mapped and monitored, if needed, water treatment should be provided or the wells closed.

ACKNOWLEDGEMENTS

The author thanks the supervisor Dr. Enn Karro for help and guidance, Dr. Leho Ainsaar for critical comments to the manuscript. I am grateful to Tuija Vaahtojärvi and Antti Vourinen for assistance with the laboratory work. The leaching experiments were performed in the Department of Geology, University of Helsinki. The preparation of this thesis was financially supported by Grant No. 5683 from the Estonian Science Foundation.

REFERENCES

- ADA, 1999. Joogivee fluoriidimine – odav ja efektiivne. *Hammas* 2, 13-15.
- Adriano, D.C. 1986. *Trace Elements in the Terrestrial Environment*. New York: Springer-Verlag. 533 p.
- Agrawal, V. & Vaish, A. 1998. Fluoride distribution in groundwater in India: sources, effects and prevention. In Weaver, T.R. & Lawrence, C.R. (eds.) *Proceedings. Groundwater: Sustainable Solutions*. IAH. Int. Groundwater Conf. 8-13 Feb., 1998. Univ. of Melbourne, Melbourne, Australia, 597-599.
- Alarcón-Herrera, M.T., Martín-Domínguez, I.R., Trejo-Vázquez, R. & Rodríguez-Dozal, S. 2001. Well water fluoride, dental fluorosis, and bone fractures in the Guadiana Valley of Mexico. *Fluoride* Vol. 34 (2), 139-149.
- Altosaar, T., Itra, A., Kreegimäe, P.H., Nõmmela, R., Runnel, R., Russak, S., Saag, M., Sirkel-Sutt, M. & Soo, T. 1998. *Laste suuhaiguste ennetamine*. BürooDisain OÜ, Tartu, 79 lk.
- Arnesen, A.K.M., Abrahamsen, G., Sandvik, G. & Krogstad, T. 1995. Aluminium-smelters and fluoride pollution of soil and soil solution in Norway. *The Science of the Total Environment* 163, 39-53.
- Barclay, J., Carroll, M., Houghton, B. & Wilson, C. 1996. Pre-eruptive volatile content and degassing history of an evolving peralkaline volcano. *Journal of Volcanology and Geothermal Research* 74, 75-87.
- Bergström, S.M., Huff, W.D., Kolata, D.R. & Bauert, H. 1995. Nomenclature, stratigraphy, chemical fingerprinting and areal distribution of some Middle Ordovician K-bentonites in Baltoscandia. *GFF* 117, 1-13.
- Biermans, V. & Baert L. 1977. Selective extraction of the amorphous Al, Fe and Si oxides using an alkaline tiron solution. *Clay Minerals* 12, 127-135.
- Carpenter, R. 1969. Factors controlling the marine geochemistry of fluorine. *Geochimica et Cosmochimica Acta* 33, 1153-1167.
- Correns, C.W. 1956. The geochemistry of the halogens, pages 181-233. In Ahrens, L.H., Rankama, K. & Runcor, S.K. *Physics and Chemistry of the Earth*. 1. Pergamon Press, London. 317 p.
- Council directive 98/83/EC. 1998. On the quality of water intended for human consumption. *Official Journal of the European Communities* L 330, 32-54.
- Cronin, S.J., Neall, V.E., Lecointre, J.A., Hedley, M.J. & Loganathan, P. 2003. Environmental hazards of fluoride in volcanic ash: a case study from Ruapehu volcano, New Zealand. *Journal of Volcanology and Geothermal Research* 121, 271-291.
- Dean, H.T. 1934. Classification of mottled enamel diagnosis. *Journal of the American Dental Association* 21, 1421-1426.
- Englund, J.O. & Myhrstad, J.A. 1980. Groundwater chemistry of some selected areas in southeastern Norway. *Nordic Hydrology* 11, 33-54.
- Fleischer, M., Forbes, R.M., Harris, R.C., Krook, L. & Kubota, J. 1974. Fluorine. Pages 22-25 in *Geochemistry and the Environment*. Vol. I. The relation of selected trace elements to health and disease. Washington, D.C.: National Academy of Science. 113 p.

- Flühler, H., Polomski, J. & Blaser, P. 1982. Retention and movement of fluoride in soil. *Journal of Environmental Quality* 11 (3), 461-468.
- Haidouti, C. 1995. Effects of fluoride pollution on the mobilization and leaching of aluminium in soils. *The Science of the total Environment* 166, 157-160.
- Hamilton, M. 1992. Water Fluoridation: A Risk Assessment Perspective. *Journal of Environmental Health* 54 (6), 27-32.
- Handa, B.K. 1975. Geochemistry and genesis of fluoride-containing ground waters in India. *Ground Water* 13, 275-281.
- Helgeson, H.C. 1969. Thermodynamics of hydrothermal systems at elevated temperatures and pressures. *American Journal of Science* 267, 279-804.
- Hem, J. 1989. Study and interpretation of the chemical characteristics of natural water. U.S. Geological Survey Water-Supply Paper 2254. 263 p.
- Heinsalu, Ü. 1995. Aluspõhja vesi. Rmt.: Raukas, A. (toim.) *Eesti Loodus*. Valgus, Tallinn, 305-311.
- Hitchon, B. 1995. Fluorine in formation waters, Alberta Basin, Canada. *Applied Geochemistry* 10, 357-367.
- Hopkins, D.M. 1977. An improved ion-selective electrode method for the rapid determination of fluorine in rocks and soils. U.S. Geological Survey, *Journal of Research* 5 (5), 589-593.
- Horn, S. & Schmincke, H.U. 2000. Volatile emission during the eruption of Baitoushan Volcano (China/North Korea) ca. 969 AD. *Bulletin of Volcanology* 61, 537-555.
- Hübner, M. 1969. Geochemische Interpretation von Fluorid/ Hydroxide Austauschversuchen an Tonmineralen. *Ber. Deutsch. Ges. geol. Wiss. B. Miner. Lagerstättenforsch.* 14, 5.
- Jacks, G. 1973. Geochemistry of fluoride in groundwater. *Swedish Dental Journal* 66, 211-215.
- Joogivee kvaliteedi- ja kontrollinõuded ning analüüsimeetodid. 2001. SOMm RTL 100, 1369.
- Jürgenson, E. 1958. Metabentoniidid Eesti NSV-s. ENSV Teaduste Akadeemia Geoloogia Instituudi uurimused, II, 73-85.
- Jürgenson, E. 1964. Siluri metabentoniidid Eesti NSV-s. Rmt: Orviku, K. jt. (toim.) *Eesti Paleosoiliste sette kivimite litoloogia*. Eesti NSV TA Geoloogia Instituut, Tallinn, 87-100.
- Karro, E., Indermitte, E., Saava, A., Haamer, K. & Marandi, A. 2006. Fluoride occurrence in publicly supplied drinking water in Estonia. *Environmental Geology* DOI: 10.1007/s00254-006-0217-1.
- Karro, E. & Rosentau, A. 2005. Fluoride levels in the Silurian-Ordovician aquifer system of western Estonia. *Fluoride* 38 (4), 307-311.
- Kiik, V. 1973. Hambakaaries ja vee fluorisisaldus Eestis. *Eesti Loodus* 9, 538-540.
- Kiipli, E. & Kallaste, T. 1996. Geochemical characterization of some Estonian metabentonites. *Proceedings of the Estonian Academy of Sciences, Geology* 45 (2), 68-77.
- Kiipli E., Kallaste T. & Kiipli T. 2004. Metabentonites of the Pirgu stage (Ashgillian, Upper Ordovician) of the East Baltic. In: Hints, O. &

- Ainsaar, L. (eds) WOGOGOB-2004 Conference Materials. Tartu University Press, Tartu, 53-54.
- Kiipli, T. 1998. Vulkanogeensed kihid Eesti settekivimites. Rmt: Reinsalu, E. (toim.) 60 aastat mäeinseneride õpetamist Eestis. Aastapäevakonverentsi ettekannete teesid ja artiklid. Tallinna Tehnikaülikooli Mäeinstituut, Tallinn, 9-12.
- Kiipli T., Kiipli E. & Kallaste T. 1997. Metabentonite composition related to sedimentary facies in the Lower Silurian of Estonia. *Proceedings of the Estonian Academy of Sciences, Geology* 46 (2), 93–104.
- Kiipli, T., Männik, P., Batchelor, R.A., Kiipli, E., Kallaste, T. & Perens, H. 2001. Correlation of Telychian (Silurian) altered volcanic ash beds in Estonia, Sweden and Norway. *Norsk Geologisk Tidsskrift* 81, 179-194.
- Koritnig, S. 1972. Fluorine 9B-9O. In Wedepohl, K.H. (ed.) *Handbook of Geochemistry*. Vol. II/1. (published in 1978). Berlin, Heidelberg, New York: Springer-Verlag.
- Kuik, L. 1964. Mikroelementide fluori- ja joodisisaldusest Eesti põhjavetes. Väitekiri farmaatsiakandidaadi kraadi taotlemiseks. Tallinn, 192 lk.
- Lahermo, P. & Backman, B. 2000. The occurrence and geochemistry of fluorides with special reference to natural waters in Finland. Report of Investigation 149. Geological Survey of Finland, Espoo, 40 p.
- Lahermo, P., Sandström, H. & Malisa, E. 1991. The occurrence and geochemistry of fluorides in natural waters in Finland and East Africa with reference to their geomedical implications. *Journal of Geochemical Exploration* 41, 65-79.
- Marshall, E. 1990. The Fluoride Debate: One More Time. *Science* 247, 276-277.
- Neal, C. 1995. Aluminium speciation variations in an acidic upland stream draining the Hafren spruce forest, Plynlimon, Mid-Wales. *Journal of Hydrology* 164 (1/4), 39-51.
- Nestor, H. 1997. Silurian. In Raukas, A., Teedumäe, A. (eds.). *Geology and mineral resources of Estonia*. Estonian Academy Publishers, Tallinn, 89-106.
- Nestor, H. & Einasto, R. 1997. Ordovician and Silurian carbonate sedimentation basin. In Raukas, A., Teedumäe, A. (eds.). *Geology and mineral resources of Estonia*. Estonian Academy Publishers, Tallinn, 192-204.
- Nestor, H. & Nestor, V. 2002. Upper Llandovery to middle Wenlock (Silurian) lithostratigraphy and chitinozoan biostatigraphy in southwestern Estonia and northernmost Latvia. *Proceedings of the Estonian Academy of Sciences, Geology* 51 (2), 67-87.
- Nestor, V. 1994. Early Silurian chitinozoans of Estonia and North Latvia. Estonian Academy Publishers, Tallinn, 163 p.
- Nicholson, K. 1983. Fluorine determination in geochemistry: errors in the electrode method of analysis. *Chemical Geology* 38, 1-22.
- Óskarsson, N. 1980. The interaction between volcanic gases and tephra, fluorine adhering to tephra of the 1970 Hekla eruption. *Journal of Volcanology and Geothermal Research* 8, 251-266.
- Perens, R. & Vallner, L. 1997. Water-bearing formation. In Raukas, A., Teedumäe, A. (eds.). *Geology and mineral resources of Estonia*. Estonian Academy Publishers, Tallinn, 137-145.

- Perens, R., Savva, V., Lelgus, M. & Parm, T. 2001. The hydrogeochemical atlas of Estonia (CD version). Geological Survey of Estonia, Tallinn.
- Raukas, A. & Teedumäe, A., 1997. Geology and mineral resources of Estonia. Estonian Academy Publishers, Tallinn, 436 p.
- Rõõmusoks, A. 1983. Eesti aluspõhja geoloogia. Valgus, Tallinn, 58-66.
- Russak, S. 1996. Hambakaarise profülaktika. Tartu Ülikool, Stomatoloogia kliinik, Tartu, 48 lk.
- Russak, S., Indermitte, E. & Saava, A. 2002. Dental fluorosis and caries among Tartu children in relation to drinking water fluoride content. Yearbook of the Estonian Anthropometric Register, 78-192.
- Saava, A. 1998. Health hazards due to drinking water. Proceedings of the Latvian Academy of Sciences, Section B, 52, 162-167.
- Saava, A., Uibo, M. & Ratnik, V. 1973. Mikroelementide sisaldus Eesti vees ja nende osa kohalikus patoloogias. Eesti Loodus 10, 606-608.
- Saxena, V.K. & Ahmed, S. 2001. Dissolution of fluoride in groundwater: a water-rock interaction study. Environmental Geology 40 (9), 1084-1087.
- Seelye, F.T. & Rafter, T.A. 1950. Low-temperature decomposition of rocks, ores and minerals by sodium peroxide using platinum vessels. Nature 4191, 317.
- Skinner, B.J. & Turekian, K.K. 1973. Man and the ocean. Englewood Cliffs: Prentice-Hall Inc.
- Smith, D.B., Zielinski, R.A. & Rose, W.A. 1982. Leachability of uranium and other elements from freshly erupted volcanic ash. Journal of Volcanology and Geothermal Research 13, 1-30.
- Smyshlyaev, S.I. & Edeleva, N.P. 1962. Determination of the solubility of minerals, I, Solubility product of fluorite. Izv. Vysshikh Uchebnyh Zavedenij Khimija i Khim. Tekhnologija 5 (6), 871-874.
- Turekian, K.K. & Wedepohl, K.H. 1961. Distribution of the elements in some major units of the Earth's crust. Geological Society of America, Bulletin 71 (2), 175-191.
- Viiding, H. 1995. Pealiskord ja selle ehitus. Rmt: Raukas, A. (koost.). Eesti Loodus. Valgus, Tallinn, 51-54.
- Vingisaar, P., Gulova, H., Kiipli, T. & Taalmann, V. 1981. Distribution of microcompounds in the Estonian Ordovician and Silurian carbonate rocks. Proceedings of the Estonian Academy of Sciences, Geology 30, 106-109.
- Vingisaar, P., Oraspõld, A., Einasto, R. & Jürgenson, E. 1965. Karbonaatkivimite ühtne klassifikatsioon ja legend. 50 p.
- Wedepohl, K.H. 1995. The composition of the continental crust. Geochimica et Cosmochimica Acta 59 (7), 1217-1232.
- Weinstein, P. & Cook, A. 2005. Volcanic emissions and health. In Selinus, O., Alloway, B., Centeno, J.A., Finkelman, R.B., Fuge, R., Lindh, U. & Smedley, P. Essentials of medical geology. Elsevier Academic Press p. 203-226.
- Wenzel, W.W. & Blum, W.E.H. 1992. Fluorine speciation and mobility in F-contaminated soils. Soil Science 153 (5), 357-364.
- WHO, 1984. Fluorine and fluorides, Environmental Health Criteria 36, IPCS International Programme on Chemical Safety, World Health Organization, Geneva.

- WHO, 2004. Guidelines for drinking-water quality. 3rd ed. World Health Organisation, Geneva.
- Yong, L. & Hua, Z.W. 1991. Environmental characteristics of regional groundwater in relation to fluoride poisoning in North China. *Environmental Geology and Water Science* 18 (1), 3-10.

FLUORIIDI HÜDROGEOKEEMIA JA GEOLOOGILISED ALLIKAD SILURI- ORDOVIITSIUMI VEEKOMPLEKSIS

Kokkuvõte

Joogiveele fluoriidi osas kehtestatud nõuetele mittevastavaks ning seega probleemseimaks veekompleksiks on Siluri-Ordoviitsiumi veekompleks. Nimetatud veekompleksi piires on võimalik välja eraldada anomaalselt kõrge F^- sisaldusega ala, mis hõlmab eelkõige Pärnu- ja Läänemaa.

S-O veekompleksi piires võivad fluoriidi sisaldused kõikuda väga laiades piirides. Üldiselt võib väita, et veekompleksi rajatud sügavad puurkaevud annavad F-rikkamat põhjavett kui madalamad. Madalamad puurkaevud jäävad aktiivsema veevahetuse vöösse, kus mikroelementide sisaldus põhjavees on reeglina madalam kui aeglase veevahetusega sügavamates kivimikihtides.

Samuti eksisteerib seos puurkaevu avatud intervalli pikkuse ning fluoriidi sisalduse vahel. Puurkaevu pikk avatud intervall võimaldab erineva mineralisatsiooniga ning erineva fluoriidi sisaldusega vete segunemist ning lahjendumist, mistõttu samasuguse sügavusega puurkaevude puhul, on kõrgemad F^- -sisaldused seotud nendega, mille töötavad intervallid on lühemad.

Põhjavee keemilises koostises eksisteerib rida seaduspärasusi, mis väljenduvad hüdrogeokeemilises võõndilisuses. Veetüüpide pindalaline muutlikkus, mis on tingitud mitmetest geoloogilistest teguritest (vettjuhtivate kivimite mineraloogilis-geokeemiline koostis, lasumissügavus, hüdrodünaamiline taustsüsteem, paleogeograafilised tingimused) määrab ka vees esineva mikroelementide koosluse. Fluoriidi esinemine põhjavees on seotud mitmete makrokomponentide (Ca^{2+} , Na^+ , Cl^- , HCO_3^-) sisalduse ja keskkonna pH-ga. Seega võib põhjavee keemiline tüüp kontrollida F^- lahustumist ja väljasettimist vees.

Fluori looduslikke allikaid võib olla mitmeid. Põhjavee keemiline koostis sõltub veega kontaktis oleva kivimi ning selle lõhetäidete koostisest, seega Eestis levivate karbonaatkivimite mikroelementide, sealhulgas fluori sisaldus on seotud kivimi tüübi, selle savikuse ja dolomiidistumise astmega. Purustatud kivimiproovide (lubjakivi, dolomiit, K-bentoniit) leostamiskatsed tõestavad K-bentoniidi fluori-rikkust, mistõttu üheks võimalikuks kõrgendatud F^- allikaks põhjavees võib lugeda vulkaanilist materjali.

Arvestades F^- toksilisust, tuleks süvendatult ja komplekselt uurida S-O põhjaveekompleksi kui Lääne-Eesti peamise joogiveeallika vee ning veekompleksi moodustavate kivimite keemilist koostist. Vältimaks suuri kulutusi fluoriidi ärastamiseks toorveest, tuleks enne uue puurkaevu rajamist uurida piirkonna hüdrogeoloogilisi ning geoloogilisi tingimusi.

APPENDIX